

Revisiting the genetic diversity and population structure of the endangered Green Sea Turtle (*Chelonia mydas*) breeding populations in the Xisha (Paracel) Islands, South China Sea

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Abstract:

The Green Sea Turtle (*Chelonia mydas*) is an umbrella species in the South China Sea, a Chinese national first-level protected wild animal, and the only sea turtle that nests in Chinese waters. The largest *C. mydas* nesting ground is distributed in the Xisha (Paracel) Islands, which plays a vital role in the survival of sea turtle populations in China. This study reveals the genetic diversity and population structure of the breeding population of *C. mydas* on the Xisha Islands using three mitochondrial markers. A total of 15 D-loop, 5 Cytochrome b (Cyt b), and 7 Cytochrome C Oxidase subunit I (COI) haplotypes were identified in the breeding population of *C. mydas* on the Xisha Islands. D-loop haplotypes are distributed in clades III, IV, and VIII of the *C. mydas* mitochondrial control region. Clade IV is first clade to be discovered in this

C. mydas population, and five D-loop haplotypes were also newly identified. The haplotype and nucleotide diversity were calculated for each marker: D-loop (0.415 haplotype diversity, 0.00204 nucleotide diversity), Cyt b (0.140, 0.00038) and COI (0.308, 0.00083). The average genetic distance (p) of each molecular marker was less than 0.01. Neutral detection and nucleotide mismatch analysis suggested that the breeding population of *C. mydas* in the Xisha Islands did not experience a population expansion event in recent history. It is recommended that a sea turtle protection area be established in the Xisha Islands area to strengthen protection and effectively protect the uniqueness and sustainability of the breeding population of *C. mydas* in the South China Sea.

Keywords: genetic diversity, green sea turtle, mitochondrial DNA, Xisha (Paracel) Islands.

Introduction

Sea turtles are large marine migratory reptiles, which are widely distributed in the Pacific, Indian, and Atlantic Ocean warm water (Hirth, 1997). As marine flagship species, sea turtles are vital for maintaining the health of marine ecosystems and are also important indicator species for marine environmental monitoring (Bouchard & Bjorndal, 2000; Hamann et al., 2010). Five sea turtle species inhabit China seas, and over 90% of the population is distributed in the South China Sea (Wang, 1993; Jiang et al., 2000; Mou et al., 2013). However, sea turtle populations have decreased sharply in China owing to coastal zone development (Chan et al., 2007), illegal commercial trade (Lam et al., 2011), marine pollution, and climate change (Wabnitz et al., 2018).

As a result, all five sea turtle species have been upgraded from level II to level I on the “List of Wildlife under Special State Protection” of China (Lin et al., 2021).

The Green Sea Turtle (*Chelonia mydas*) is the only sea turtle species that lays eggs in China seas (Wang et al., 2019). Its nesting grounds were once widely distributed in China's southern coastal areas in Hainan, Guangxi, Guangdong, and Fujian Provinces. However, most of these nesting grounds have disappeared in the last century, including in the Huidong Sea Turtle National Nature Reserve, which has had no recorded nests since 2018 (Lin et al., 2021). The Xisha or Paracel Islands (hereafter referred to as the Xisha Islands) are currently the largest nesting grounds for *C. mydas* in China, with more than 100 nests recorded annually (Wang et al., 2019; Jia et al., 2019). Recent studies have found that *C. mydas* from the Xisha Islands rookery represent a new geographic population with unique haplotypes and a new conservation management unit with a high conservation value (Gaillard et al., 2021; Song et al., 2022).

Genetic diversity is an important feature of a species, reflecting its potential to adapt to environmental change (Ekanayake et al., 2017). Sea turtles are long-lived organisms, and their life histories are marked by ontogenetic habitat shifts and large-scale migrations (Bowen & Karl, 2007). Therefore, accurately and comprehensively understanding sea turtle genetic diversity will help develop effective conservation strategies. Due to its simple structure and relatively rapid evolution rate, mitochondrial DNA (mtDNA) molecular markers are the commonly used genetic markers for assessing sea turtle population structure, genetic diversity, and

phylogeography (Naro-Maciel et al., 2008; Guo, Wang & Liu, 2009; Leroux et al., 2012; Vargas et al., 2016; Yang, 2015). Yang (2015) and Wei (2016) both used mitochondrial genes to do a preliminary study of the genetic diversity of *C. mydas* in the South China Sea, but their samples were mainly juvenile sea turtles from eggs that were incubated artificially or from *C. mydas* confiscated by the coast guard. Unclear sample collection site may not reflect the level of genetic diversity in the wild breeding population of *C. mydas* in South China Sea. Gaillard et al. (2021) and Song et al. (2022) also studied breeding populations of *C. mydas* in the Xisha Islands, but their sample sizes were small ($n = 16$ or $n = 13$), limiting the application of the data to resource management and conservation of *C. mydas* breeding populations in this area.

From 2017 to 2021, a total of 72 individual samples were collected from the *C. mydas* nesting grounds in the Xisha Islands. This study aimed to (1) comprehensively evaluate the genetic diversity of the breeding population of *C. mydas* in the Xisha Islands using the three mtDNA markers and understand the evolutionary and adaptive potential of *C. mydas* in this region and (2) analyze the genetic structure and historical dynamics of the population. This work will provide a scientific basis for management activities of the *C. mydas* breeding population inhabiting the Xisha Islands.

Materials & Methods

Sampling

During the breeding season of *C. mydas* in the Xisha Islands of the South China Sea, biopsy punch sampling methods were used to collect hind limbs skin samples from female turtles that came ashore to lay eggs. Biopsy sites were disinfected with

iodophor (LIRCON, Shandong, China) before and after sampling. After the female turtle lays eggs, we used the scanner (RBC-S03, Raybaca IOT Technology, Anhui, China) to identify whether the female turtle has Passive Integrated Transponder (PIT) markers (RBC-Z00, Raybaca IOT Technology, Anhui, China), and make chip markers and samples for the newly discovered female turtle, so as to ensure that the sampling of female turtles was not repeated. For the nest corresponding to the identified female turtle, we only took a sample from the female turtle. If egg-laying female turtles were not identified for nests, one sample was taken from each nest, either as a forelimb skin of a hatchling or from embryos that had died in the nests. Between 2017 and 2021, a total of 72 samples (Yongle Islands: 12 samples; Xuande Islands: 60 samples) were collected, and all samples were stored in 95% alcohol at -20 °C (Table S1, Fig. 1). The sample collection work was approved by the Chinese government, and this work was conducted in strict accordance with the guidelines of the Animal Research Ethics Committee of Hainan Provincial Education Centre for Ecology and Environment, Hainan Normal University (HNECEE-2012-005).

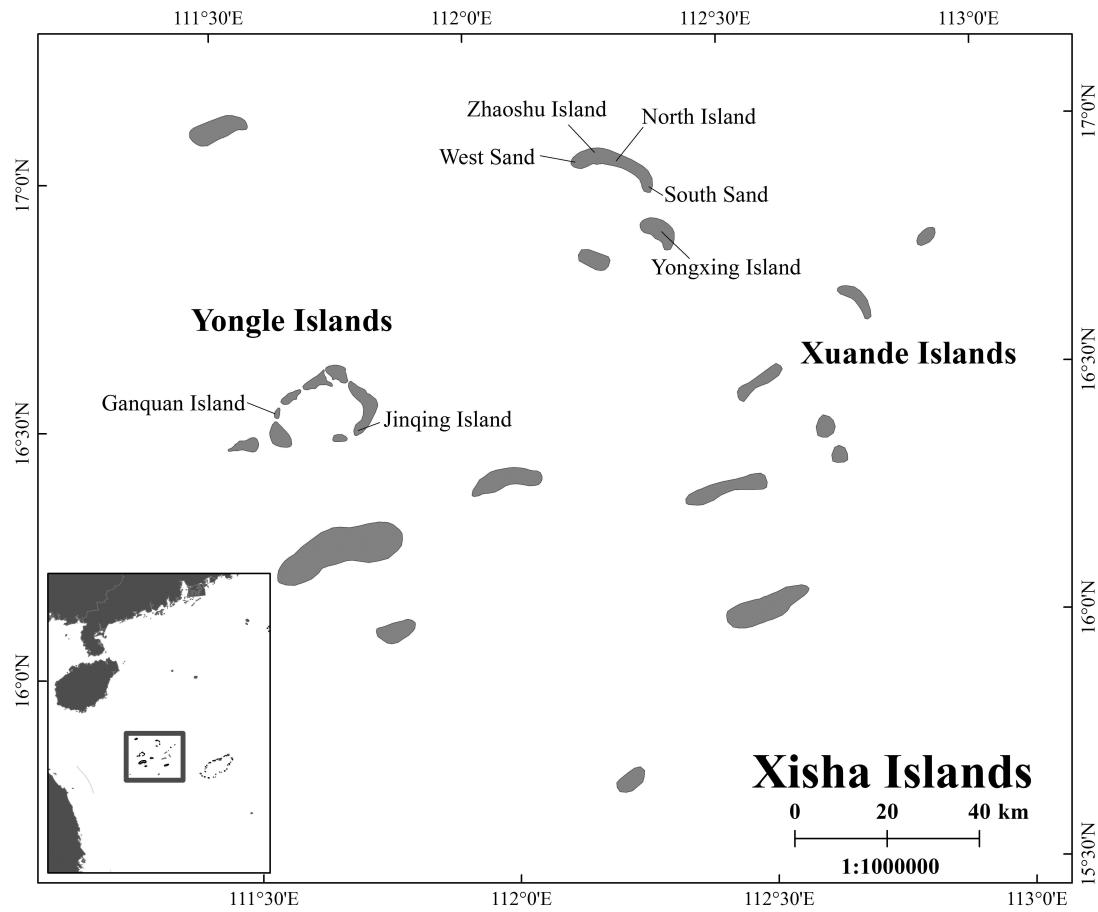


Fig.1 Map of study area and sampling points. Sampling points: West Sand; North Island; South Sand; Ganquan Island; Jinqing Island

Laboratory work

DNA was extracted from skin samples using a blood/cell/tissue genomic DNA extraction kit (DP304, Tiangen Biotech Co., Ltd., Beijing, China). DNA concentrations were all above 4 ng/μl.

Three sets of primers were used to amplify the target sequences of the control region (D-loop), cytochrome b (Cyt b), and Cytochrome C Oxidase subunit I (COI) (Table 1).

Table 1 Mitochondrial DNA primers for *C. mydas*

Marker	Primer	Sequence	Product length (bp)	Reference
D-loop	LCM15382	GCTTAACCCTAAAGCATTGG	770	(Abreu-Grobois et al., 2006)
	H950g	AGTCTCGGATTAGGGTTG		
Cytb	Cytb - f	ATTCTTGCTGGACTTA	1140	(Yang, 2015)
	Cytb - r	TTCAATCTTGGTTACA		
COI	COI - f	TCAACCAACCACAAAGACATTG	650	(Yang, 2015)
		GCAC		
	COI - r	TAGACTTCTGGTGGCCAAAGA		
		ATCA		

PCR was performed in 50 μ l reactions, with 25 μ l of 2 \times Taq Mix (RN03001S, MonAmpTM), 2 μ l of template DNA, 2 μ l of each 10 uM primer, and ultrapure water to 50 μ l. The PCR conditions of D-loop as follows: pre-denaturation at 94 $^{\circ}$ C for 3 min; followed by 30 cycles of denaturation at 94 $^{\circ}$ C for 30 s, annealing at 51 $^{\circ}$ C for 30 s, extension at 72 $^{\circ}$ C for 60 s; and final extension at 72 $^{\circ}$ C for 3 min. The PCR conditions of Cyt b as follows: pre-denaturation at 94 $^{\circ}$ C for 80 s; followed by 30 cycles of denaturation at 94 $^{\circ}$ C for 42 s, annealing at 47 $^{\circ}$ C for 30 s, extension at 72 $^{\circ}$ C for 80 s; and final extension at 72 $^{\circ}$ C for 5 min. The PCR conditions of COI as follows: pre-denaturation at 94 $^{\circ}$ C for 80 s; followed by 30 cycles of denaturation at 94 $^{\circ}$ C for 42 s, annealing at 55 $^{\circ}$ C for 30 s, extension at 72 $^{\circ}$ C for 80 s, and final extension at 72 $^{\circ}$ C for 5 min. PCR products were checked with a 1% agarose gel and successful reactions purified and Sanger sequenced using PCR primers at Guangzhou Ige Biotechnology Co., Ltd.

Data Analysis

(1) Sequence composition analysis

D-loop, Cyt b, and COI Sequences were spliced using Sequencher 5.4.5 software. MEGA-X was used to align the sequences and calculate the number of parsimony informative sites and variable sites.

(2) Haplotype and genetic diversity Analysis

Shorter control region segments were commonly used to align and analyze population structures in *C. mydas*, in particular a 384 bp region found in the Indo-Pacific region. Accordingly, the ~753 bp sequences obtained were subsequently trimmed to 384 bp for further comparison with the available data using the standardized CmP nomenclature.

BLAST was used to compare sequences to GenBank to identify existing haplotypes or confirm new haplotypes. DnaSP5.0 software was used to detect the number of haplotypes, nucleotide, haplotype diversity, and average number of nucleotide differences.

(3) Population genetic differentiation analysis

MEGA-X was used to align the sequences and calculate the genetic distance of each haplotype. Genetic differentiation index Fst and gene flow (Nm) between two geographic populations (Xuande, Yongle) were obtained using DnaSP5.0 software based on sequences data. The software Network10.2 was used to draw the haplotype network diagram, which reflected the evolutionary relationship among haplotypes.

IQ-TREE (Trifinopoulos 2016) was used to determine the best-fit model of nucleotide substitutions. The software BEAST was used to perform Bayesian phylogenetic reconstructions. The MEGA-X software was used to construct a

phylogenetic tree based on the Hasegawa-Kishino-Yano model with the Maximum Likelihood method (ML), including 11 D-loop haplotypes from five evolutionary clades of *C. mydas* distributed in the Indo-Pacific control region (Jensen et al., 2019), and the *Eretmochelys imbricata* (GenBank AJ421794) as an outgroup (Table S2). The confidence of each branch of the ML tree was tested by 1000 bootstrapping repetitions, retaining the default values for other parameters.

(4) Population dynamic history analysis

For each of the two populations (Xuande, Yongle), Fu's Fs and Tajima's D neutrality tests and mismatch distribution analysis were carried out using DnaSP5.0 software.

Results

Sequence composition

The number of new sequences obtained from the Xisha Islands breeding population for each marker is as follows: 72 D-loop (753 bp length), 69 Cyt b (1052 bp), and 71 COI (511 bp). The parsimony informative sites and variable sites were calculated for each marker: D-loop (7 parsimony informative sites, 38 variable sites), Cyt b (2, 12), and COI (4, 6).

Haplotype and genetic diversity

A total of 12 D-loop, 5 Cyt b, and 7 COI haplotypes were identified from the 72 samples collected from the nesting grounds of the Xisha Islands (Table 2). The haplotype diversity (h) of each marker ranged from 0.140 to 0.415, while the nucleotide diversity (n) ranged from 0.00038 to 0.00204 (Table 3). Based on the

cut-off point of haplotype diversity (0.5) and nucleotide diversity (0.005) set by Grant et al. (1998), the breeding population of *C. mydas* in the Xisha Islands showed low haplotype diversity and nucleotide diversity. The haplotype diversity of the D-loop was 0.415 ± 0.07 , slightly lower than the mean value found for the Indo-Pacific *C. mydas* management units (mean 0.472 ± 0.07). The nucleotide diversity was 0.00204 ± 0.00 , lower than the Indo-Pacific *C. mydas* mean of 0.0088 ± 0.00 (Table 4) (Dethmers et al., 2006; Cheng et al., 2008; Nishizawa et al., 2011; Jensen et al., 2016).

Table 2 D-loop Haplotypes and frequencies of *C. mydas* from the Xisha Islands

haplotype	samples	Frequency(%)	samples	frequency(%)
	Xuande Islands		Yongle Islands	
D-loop	60		12	
CmP18.1	1	1.7	1	8.3
CmP19.1	46	76.7	9	75
CmP47.1	0	0	1	8.3
CmP49.1	1	1.7	1	8.3
CmP57.1	1	1.7	0	0
CmP75.1	1	1.7	0	0
CmP154.1	4	6.7	0	0
CmP250.1*	1	1.7	0	0
CmP251.1*	1	1.7	0	0
CmP252.1*	2	3.3	0	0
CmP253.1*	1	1.7	0	0
CmP254.1*	1	1.7	0	0
Cytb	57		12	
CMB1*	53	93	11	91.7
CMB2*	2	3.5	0	0
CMB3*	0	0	1	8.3
CMB4*	1	1.8	0	0
CMB5*	1	1.8	0	0
COI	59		12	
CMC1	50	84.7	9	75
CMC2*	2	3.4	0	0
CMC3*	1	1.7	1	8.3
CMC4*	1	1.7	0	0

CMC5	3	5.1	1	8.3
CMC6*	2	3.4	0	0
CMC7	0	0	1	8.3

* Represents the haplotype newly identified in this study.

Table 3 The genetic diversity parameters and the haplotype average genetic distance of *C. mydas*

individuals of this study, separated by island group

Marker	group	# samples	# haplotypes	Haplotype diversity	Nucleotide diversity	Average nucleotide difference number (<i>k</i>)	Average genetic distance (<i>p</i>)	variable sites	informative sites
D-loop	Xuande	60	11	0.411 ± 0.080	0.00099 ± 0.000	0.740			
	Yongle	12	4	0.455 ± 0.170	0.00733 ± 0.003	5.470			
	Total	72	12	0.415 ± 0.074	0.00204 ± 0.002	1.522	0.010	38	7
Cytb	Xuande	57	4	0.136 ± 0.061	0.00030 ± 0.000	0.313			
	Yongle	12	2	0.167 ± 0.134	0.00079 ± 0.000	0.833			
	Total	69	5	0.140 ± 0.057	0.00038 ± 0.000	0.404	0.0046	12	2
COI	Xuande	59	6	0.281 ± 0.076	0.00070 ± 0.000	0.359			
	Yongle	12	4	0.455 ± 0.170	0.00146 ± 0.001	0.742			
	Total	71	7	0.308 ± 0.071	0.00083 ± 0.000	0.426	0.0040	6	4

Table 4 Comparison of genetic diversity of D-loop marker between the Xisha Islands and other locations in the Indo-Pacific region

Country	management unit	# haplotypes	Haplotype diversity (<i>h</i>) \pm SD	nucleotide diversity (<i>n</i>) \pm SD	# samples
China	Western Taiwan ^C	3	0.483 ± 0.06	0.028 ± 0.01	40
China	Eastern Taiwan ^C	1	0.000 ± 0.00	0.000 ± 0.00	14
China	Xisha Islands	12	0.415 ± 0.07	0.002 ± 0.00	72
Indonesia	Western Java ^A	2	0.485 ± 0.06	0.001 ± 0.00	22
Indonesia	Aru ^A	2	0.071 ± 0.07	0.004 ± 0.00	28
Indonesia	Sangalaki ^D	5	0.780 ± 0.16	0.008 ± 0.03	29
Malaysia	Sipadan ^A	7	0.630 ± 0.07	0.005 ± 0.00	98
Malaysia	Peninsula Malaysia ^A	7	0.567 ± 0.10	0.008 ± 0.00	29
Malaysia	Sarawak ^A	3	0.450 ± 0.11	0.009 ± 0.00	22

Malaysia and Philippines	Turtle Islands ^A	3	0.369 ± 0.06	0.001 ± 0.00	66
Australia	Cocos Keeling Island ^A	2	0.199 ± 0.11	0.011 ± 0.01	19
Australia	Ashmore Reef ^A	7	0.670 ± 0.04	0.030 ± 0.02	44
Australia	Scott Reef and Browse Island ^A	5	0.498 ± 0.06	0.007 ± 0.00	64
Australia	North West Shelf ^A	8	0.433 ± 0.07	0.004 ± 0.00	76
Australia	Cobourg Peninsula ^A	5	0.574 ± 0.08	0.003 ± 0.00	37
Australia	Gulf of Carpentaria ^A	6	0.637 ± 0.02	0.004 ± 0.00	127
Japan	Ogasawara ^B	13	0.706 ± 0.04	0.018 ± 0.01	103
	Mean	5.35	0.472 ± 0.07	0.0088 ± 0.00	52.35

^A Jensen et al., (2016).

^B Nishizawa et al., (2011).

^C Cheng et al., (2008).

^D Dethmers et al., (2006).

Among the 12 D-loop haplotypes, seven were found in previous studies (CmP18.1, CmP19.1, CmP47.1, CmP49.1, CmP57.1, CmP75.1 and CmP154.1) (Gaillard et al., 2021; Song et al., 2022), while five were newly discovered (CmP250.1, CmP251.1, CmP252.1, CmP253.1 and CmP254.1, with GenBank accession numbers OK284742, OK324138, OP320709, OP320710, and OP320711, respectively), and they were all found in the Xuande Islands. The most common haplotype in the samples was CmP19.1, accounting for 76.4% of the total samples from the nesting grounds of the Xisha Islands, followed by CmP154.1, accounting for 5.6% and CmP18.1, CmP49.1, and CmP252.1, each accounting for 2.8%. All five Cyt b haplotypes were newly discovered (CMB1–CMB5) and haplotype CMB1 was the most common, with a frequency of 92.75%. For the seven COI haplotypes, four (CMC2, CMC3, CMC4, and CMC6) were newly discovered in this study. Among

them, haplotype CMC1 was the most common, with a frequency of 83.1%, and the other haplotypes were relatively rare.

Population genetic differentiation

Of the breeding population of *C. mydas* in the Xisha Islands, the average genetic distances of the D-loop, Cyt b, and COI haplotypes were 0.010, 0.0046, and 0.0040, respectively (Table 3). The genetic distance between D-loop haplotype CmP47.1 and the other 11 haplotypes was the largest (0.04), with most of the genetic distances between 0.001 and 0.008. The genetic distances between the haplotypes of the Cyt b and COI genes ranged from 0.001 to 0.009 and 0.002 to 0.006, respectively.

The genetic differentiation coefficient Fst between the two geographic groups of *C. mydas* in Xuande and Yongle Islands was -0.00690–0.01303, and the gene flow coefficient (Nm) was -36.46–114.37 (Table 5), indicating that there was no genetic differentiation between the two geographic populations and gene exchange was frequent.

Table 5 Fst values and gene flow between two geographical groups of *C. mydas* in Xuande Islands and Yongle Islands

Marker	Fst	gene flow (Nm)
D-loop	-0.00690	-36.46
Cyt b	0.00218	114.37
COI	0.01303	18.93

Combining the seven 384 bp D-loop haplotypes (CmP18, CmP19, CmP20, CmP49, CmP54, CmP75, and CmP244) found in Gaillard et al. (2021) and Song et al.

(2022), there are, in total, 15 D-loop haplotypes in the breeding population of *C. mydas* in the Xisha Islands (Table 6).

With reference to the clades studied by Jensen et al. (2019), the Xisha Islands D-loop haplotypes belong to three evolutionary branches: three haplotypes (CmP20, CmP54, and CmP244) belong to evolutionary clade III, one haplotype (CmP47) belongs to clade IV, and the other 11 haplotypes belong to clade VIII (Table 6, Figure 2). The Bayesian tree has the same topology as the Maximum Likelihood tree.

Table 6 The clade, haplotypes (384 bp control region) and individuals number of *C. mydas* in Xisha Islands, South China Sea (Clade reference Jensen et al., 2019)

Clade	haplotype	# individuals	reference
III	CmP20	1	Song et al. (2022)
	CmP54	1	Gaillard et al. (2021)
	CmP244	1	Song et al. (2022)
IV	CmP47	1	This study
VIII	CmP18	4	Song et al. (2022)
			Gaillard et al. (2021)
	CmP19	73	Song et al. (2022)
			Gaillard et al. (2021)
			This study
	CmP49	6	Gaillard et al. (2021)
			This study
	CmP57	1	This study
	CmP75#	1	This study
	CmP154	4	This study
Total	CmP243#	2	Song et al. (2022)
	CmP250	1	This study
	CmP251	1	This study
	CmP252	2	This study
	CmP253	1	This study
	CmP254	1	This study
		101	

For the same haplotype, the CmP243 haplotype was actually CmP75 haplotype.

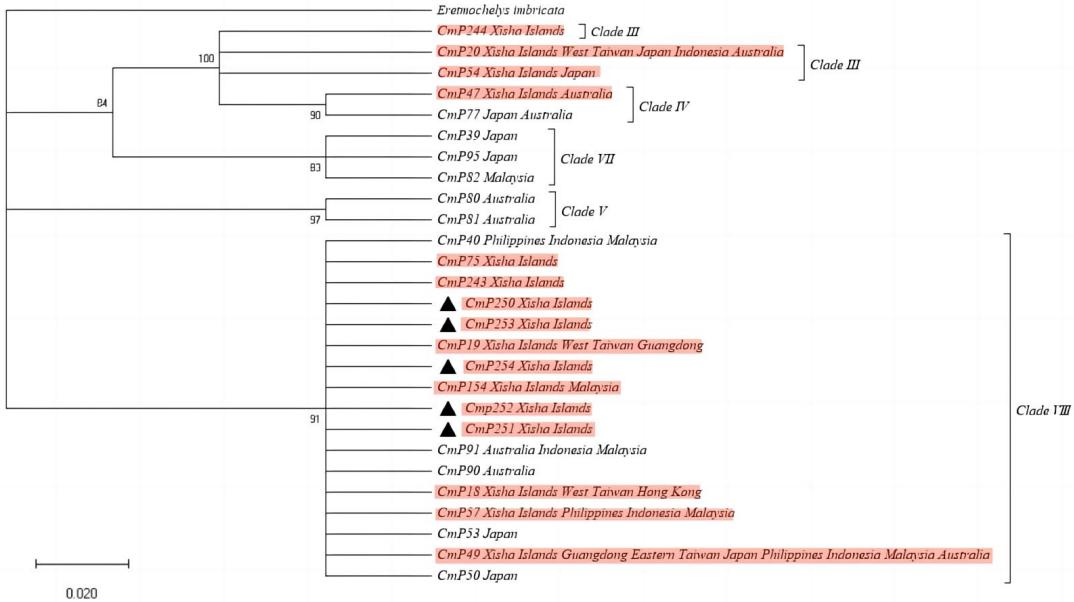


Fig. 2 ML tree of 26 D-loop haplotypes (384 bp control region) from rookeries of *Chelonia mydas* in the Indo-Pacific and Japan regions, including 15 haplotypes from the Xisha Islands (red text). Numbers above branches represent bootstrap values based on 1,000 replicates. The clades used here follow Jensen et al. (2019). ▲ indicates the newly discovered haplotype in this study.

The haplotype network diagrams (Figures 3–5) reflect the evolutionary relationship among haplotypes of various genes of *C. mydas* in the Xisha Islands. As shown in Figure 3, the D-loop haplotype network is mainly divided into two clusters (A and B), a result largely consistent with the ML tree topology. The haplotypes in cluster A are all distributed in clade VIII (Jensen et al., 2019), and the most numerous haplotype CmP19.1 is the central haplotype. The haplotypes in cluster B are distributed in clades III and IV (haplotypes CmP20.1, CmP54.1, CmP244.1, and CmP47.1). The haplotype network diagram of Cyt b haplotypes shows radial divergence, with the most frequent haplotype, CMB1, as the central haplotype. The haplotype network of COI haplotypes was also radial, with the most frequent haplotype, CMC1, as the central haplotype.

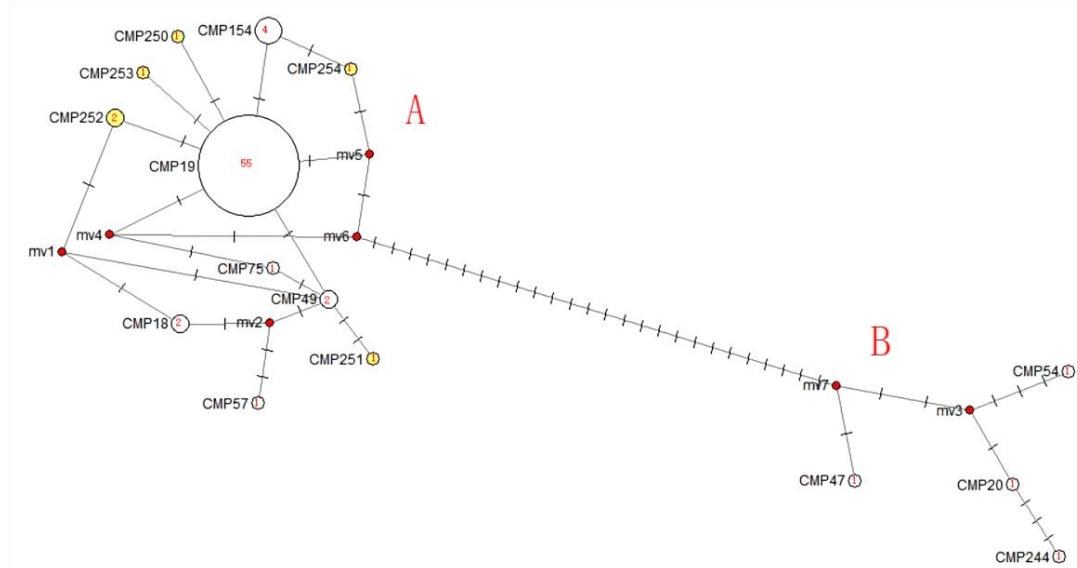


Fig. 3 The haplotype network diagram showing the relationships between the 15 D-loop haplotypes among the *C. mydas* in Xisha Island. Hatch mark represents the number of mutational steps between pairs of haplotypes. The red dot represents an unsampled intermediate haplotype connecting sampled haplotypes. The yellow circles represent the haplotypes newly identified in this study. Numbers in the circles indicated sample size of each haplotype.

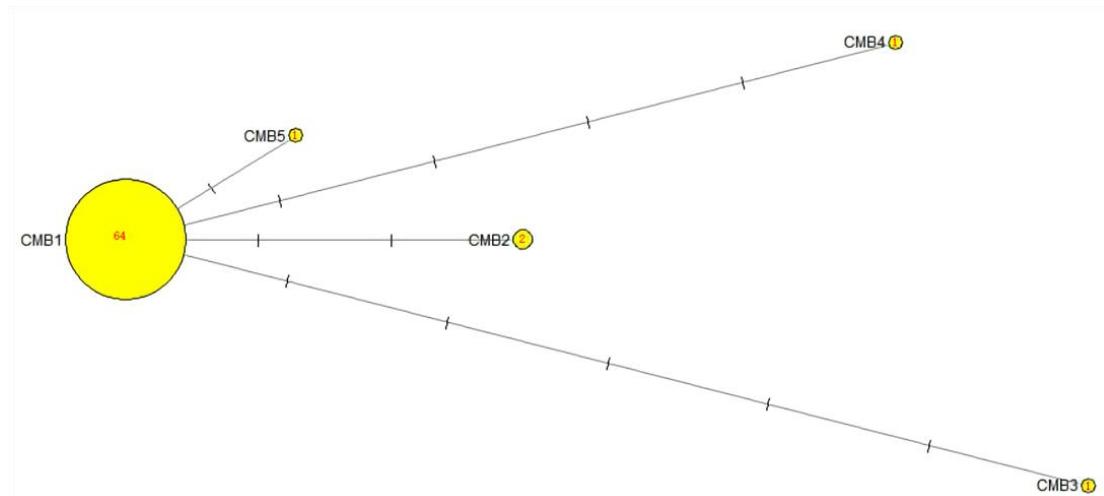


Fig. 4 The haplotype network diagram showing the relationships between the 5 Cyt b haplotypes among the *C. mydas* in Xisha Island.

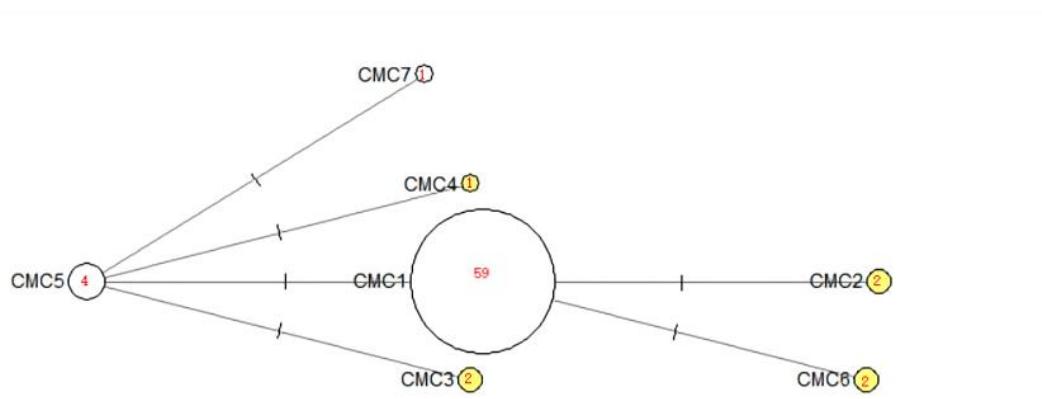


Fig. 5 The haplotype network diagram showing the relationships between the 7 COI haplotypes among the *C. mydas* in Xisha Island.

Population Dynamic History

Tajima's D and Fu's Fs values of each marker are shown in Table 7. Figures 6–8 show a distribution diagram of nucleotide mismatches of three markers, all showing a single peak. Combining the results from Tajima's D value, Fu's Fs value and the nucleotide mismatch distribution map of each marker, we infer that the breeding population of *C. mydas* in the Xisha Islands did not experience a population expansion.

Table 7 Neutral detection parameters of *C. mydas* in Xisha Islands, means P-value < 0.01

Marker	# individuals	Neutral detection	
		Tajima's D	Fu's Fs
D-loop	72	-2.60518**	-4.361
Cytb	69	-2.35107**	-2.338
COI	71	-1.56483	-4.896

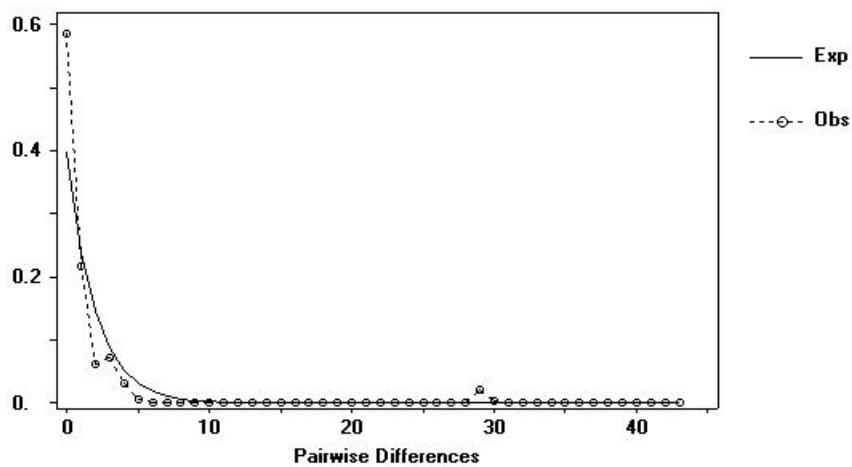


Fig. 6 Nucleotide mismatch distributions of the D-Loop markers used in this study.

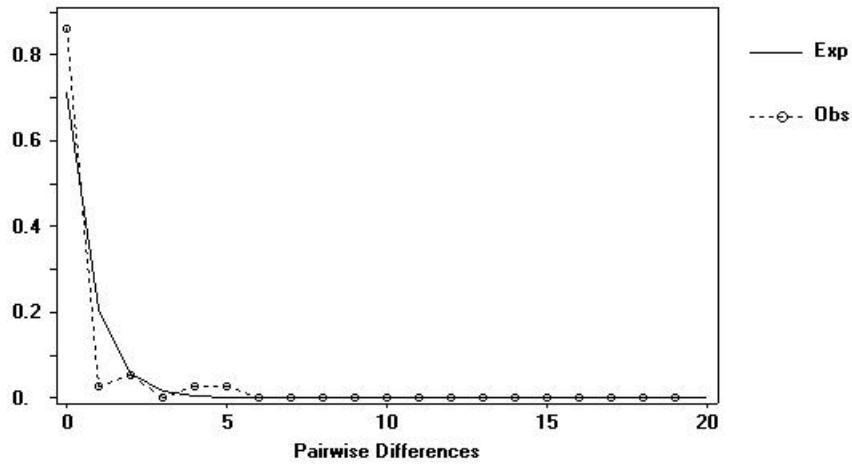


Fig. 7 Nucleotide mismatch distributions of the Cyt b markers used in this study.

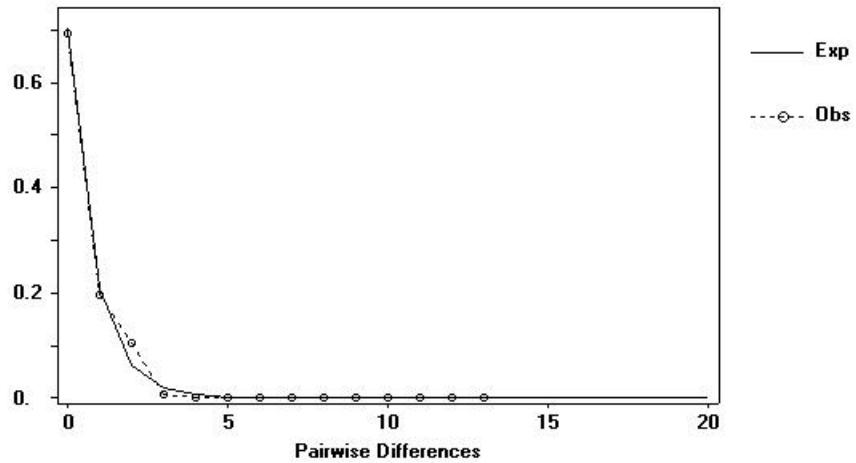


Fig. 8 Nucleotide mismatch distributions of the COI marker used in this study.

Discussion

Gaillard et al. (2021) and Song et al. (2022) found that the breeding population of *C. mydas* from Xisha Islands represent a new geographic population with the unique D-loop haplotype CmP19, which is only sporadically found in other breeding or feeding grounds in the Indo-Pacific region. When combining the haplotypes identified by Gaillard et al. (2021) and Song et al. (2022), we found that the CmP19 haplotype represented 73% (n=101) of the *C. mydas* breeding population in the Xisha islands. Therefore, our study further confirms the uniqueness of this population. We identified five new D-loop haplotypes in the Xisha islands, increasing the number of haplotypes in this population from the original 7 to 15. Additionally, according to the D-loop haplotype network (Figure 3), there are at least seven intermediary haplotypes may exist in nature, but these have not yet been sampled in this area. In this study, nine of the 15 D-loop haplotypes, three of the five Cyt b haplotypes, and two of the seven COI haplotypes represent only a single *C. mydas* sample. If more samples can be collected, a more reliable estimate of haplotype frequency can be provided. Thus, it is necessary to increase the sampling range and sample numbers in the future to find more potential haplotypes of the sea turtle population in this area.

In this study, the D-loop genetic diversity (haplotype diversity: 0.415, nucleotide diversity: 0.002) of *C. mydas* from the Xisha Islands is much smaller than that of Gaillard et al. (2021) and Song et al. (2022) (0.575, 0.009 and 0.628, 0.018, respectively). This could be due to the small sample sizes (16 and 13, respectively) in the previous two studies. Small sample sizes can greatly reduce resolution, the ability

to discriminate between populations and the power to detect rare or unique haplotypes, which may lead to discrepancies between actual and observed gene richness (Marshall & Brown, 1975). Among the 72 samples in this study, the number of D-loop haplotypes was 12, higher than the average (5.35, Table 4) of management units in the Indo-Pacific region. Although the number of samples has been greatly increased in our study, it is still small compared with other regions, such as the Galapagos Islands (Duttonet al., 2014). This is mainly because it is quite difficult to reach the Xisha islands, and *C. mydas* breeding population there is low.

Based on D-Loop haplotypes, Jensen et al. (2019) identified five evolutionary clades (clade III, IV, V, VII, and VIII) of *C. mydas* in the Indo-Pacific region, with clade VIII predominant. At present, individuals from the breeding population of the Xisha Islands were found in three of these clades (III, IV, VIII), with haplotypes in VIII dominating (73%), consistent with Jensen et al. (2019). The *C. mydas* population of the Xisha Islands shared four haplotypes with *C. mydas* from the Philippines, Malaysia, Indonesia, and Australia (CmP20, CmP47, CmP49, and CmP57). However, these four haplotypes only account for 8.9% of the total samples studied from the Xisha Islands ($n = 101$) and were relatively rare in the Xisha Islands. MtDNA shows only female-mediated gene flow, and will not actually show the contribution from males, who may be from outside the South China Sea. Therefore, further sampling, including more individuals and more nuDNA markers as well as telemetry data, will be needed to determine whether *C. mydas* in the Xisha islands are communicating frequently with *C. mydas* from adjacent areas.

A reduction in population size often leads to a loss of genetic diversity, which reduces reproduction and survival (e.g., inbreeding depression) and further reduces population sizes (Lacy, 1997; Frankham, 2005). Sea turtle populations in Chinese waters have dropped sharply due to the massive illegal trade and habitat loss (Lin et al., 2021). The exploitation and trade of turtle products was common in Southeast Asia, especially in Malaysia, Indonesia, and the Philippines, where sea turtle populations have suffered from chronic illegal hunting (Dijk & Shepherd, 2004; Stiles, 2008; Lam et al., 2011). Gaillard et al. (2021) showed that the *C. mydas* recently confiscated in Hainan likely come from the Coral Triangle, and populations in the Xisha Islands and the Sulu Sea are facing serious illegal hunting pressure. The degradation of nesting beaches has also gradually reduced the breeding space of *C. mydas* in the Xisha Islands. For example, sea turtles used to lay eggs on Yongxing Island and Zhaoshu Island (Wei, 2016; Jia et al., 2019), but, due to human activities, they no longer go ashore there. In addition, marine pollution is threatening the remaining nesting beaches and breeding populations of sea turtles in the Xisha islands. Beach debris, microplastic, and trace element pollution are widespread in the nesting grounds of *C. mydas* in the Xisha Islands, and those pollutants have potential negative impacts on the reproductive activities and embryo development of sea turtles (Zhang et al., 2020; Zhang et al., 2021a; Jian et al., 2021; Zhang et al., 2022). Evidence of global warming was also recorded in the Xisha islands, as beach temperatures have increased by 1–2 °C from 2018 to 2021 (Zhang et al., 2022). It is well-known that the sex of sea turtles is determined by the hatching temperature. The increase in hatching

temperature will increase female offspring, which has been confirmed by laparoscopic surgery performed on sea turtles from South China Sea (Yeh et al., 2021). This gender imbalance will threaten reproduction and decrease turtle populations in the future (Stewart & Dutton, 2014). The positive feedback loop between small population size and low genetic diversity, termed “the vortex effect”, may ultimately lead to extinction (Gilpin & Soulé, 1986; Hughes et al., 2008).

Green sea turtles are a migratory marine species with a wide distribution range and are highly loyal to their breeding and feeding grounds (Formia et al., 2006; Hamabata et al., 2014). Confiscated *C. mydas* in the South China Sea affect not only the breeding population of *C. mydas* in the Xisha Islands, but also other distant nesting grounds, such as the Philippines and Malaysia (Gaillard et al., 2021; Zhang et al., 2021b). Therefore, it is necessary to strengthen cross-border cooperation in monitoring and protecting sea turtle populations in the South China Sea and internationally. Furthermore, it is necessary to completely prohibit the commercial use of sea turtles as live or wildlife products and to monitor market trends and trade routes to jointly protect sea turtle populations. The Xisha Islands are also a key area for Chinese fishermen who operate in the South China Sea (Wang et al., 2019). To strengthen the protection of sea turtles, the Chinese Government issued the “Sea Turtle Conservation Action Plan (2019–2033)” (Chinese Ministry of Agriculture and Rural Affairs, 2018) and upgraded all five sea turtle species from level II to level I on the “List of Wildlife under Special State Protection” of China in 2021 (Lin et al., 2021). These policies make the prospect of sea turtle protection brighter in the South

China Sea. We recommend that a sea turtle sanctuary be established in the Xisha Islands to protect nesting and foraging sites, monitor nesting populations, and mitigate the effects of fishing and costal development on this population of *C. mydas*.

When analyzing the samples of the breeding population of *C. mydas* in the Xisha Islands, this study only analyzed the mitochondrial genes of *C. mydas*, which is relatively simple. In the later stage, it is urgent to combine nuclear genes, microsatellites, and other technologies (such as Single Nucleotide Polymorphisms analysis or Restriction Site-associated DNA sequencing) to comprehensively evaluate the genetic diversity of the breeding populations of *C. mydas* in the Xisha Islands. Concurrently, it is necessary to further explore the feeding grounds of *C. mydas*, strictly plan fishing areas, and strengthen scientific outreach to communities for the public to effectively protect China's sea turtle populations.

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Author Contributions

Meimei LI conceived and designed the experiments, performed the experiments, analyzed the data, prepared figures and/or tables, authored or reviewed drafts of the

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Ting ZHANG performed the experiments, analyzed the data, prepared figures and/or tables, authored or reviewed drafts of the paper, and approved the final draft.

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References

Abreu-Grobois FA, Horrocks JA, Formia A, Dutton PH, LeRoux R A, Velez-Zuazo X, Soares LS, Meylan AB. 2006. New mtDNA D-loop primers which work for a variety of marine turtle species may increase the resolution of mixed stock analysis. In Proceedings of the 26th Annual Symposium on Sea Turtle Biology, Book of Abstracts DOI [10.1007/s10592-005-9102-9](https://doi.org/10.1007/s10592-005-9102-9).

Andrade AL. 2011. Microplastics in the marine environment. *Marine Pollution Bulletin* 62(8): 1596-1605 DOI [10.1016/j.marpolbul.2011.05.030](https://doi.org/10.1016/j.marpolbul.2011.05.030).

Bouchard SS, Bjorndal KA. 2000. Sea Turtles as Biological Transporters of Nutrients and Energy from Marine to Terrestrial Ecosystems. *Ecology* 81(8): 2305-2313 DOI [10.1890/0012-9658\(2000\)081\[2305:STABTO\]2.0.CO;2](https://doi.org/10.1890/0012-9658(2000)081[2305:STABTO]2.0.CO;2).

Bowen BW, Karl SA. 2007. Population genetics and phylogeography of sea turtles. *Molecular Ecology* 16: 4886-4907 DOI [10.1111/j.1365-294X.2007.03542.x](https://doi.org/10.1111/j.1365-294X.2007.03542.x).

Chan SKF, Cheng IJ, Zhou T, Wang HJ, Gu HX, Song XJ. 2007. A Comprehensive Overview of the Population and Conservation Status of Sea Turtles in China. *Chelonian Conservation and Biology* 6(2): 185-198 DOI [10.2744/1071-8443\(2007\)6\[185:ACOOTP\]2.0.CO;2](https://doi.org/10.2744/1071-8443(2007)6[185:ACOOTP]2.0.CO;2).

Cheng IJ, Dutton PH, Chen CL, Chen HC, Chen YH, Shea JW. 2008. Comparison of the genetics and nesting ecology of two green sea turtle rookeries. *Journal of Zoology* 276(4):375-384 DOI [10.1111/j.1469-7998.2008.00501.x](https://doi.org/10.1111/j.1469-7998.2008.00501.x).

Chinese Ministry of Agriculture and Rural Affairs. 2018. Sea Turtle Conservation Action Plan (2019–2033) http://www.moa.gov.cn/nybgb/2019/201901/201905/t20190503_6288226.htm.

Dethmers KM, Broderick D, Moritz C, Fitzsimmons NN, Limpus CJ, Lavery S, Whiting S, Guinea M, Prince RIT, Kennett R. 2006. The genetic structure of Australasian green turtles (*Chelonia mydas*): exploring the geographical scale of genetic exchange. *Molecular Ecology* 15(13): 3931-3946 DOI [10.1111/j.1365-294X.2006.03070.x](https://doi.org/10.1111/j.1365-294X.2006.03070.x).

Dijk VPP, Shepherd CR. 2004. Shelled Out? A Snapshot of Bekko Trade in Selected Locations in Southeast Asia. TRAFFIC, Southeast Asia, Petaling Jaya, Selangor, Malaysia.

Dutton PH, Jensen MP, Frey A, LaCasella E, Balazs GH, Zárate P, Chassín-Noria O, Sarti-Martinez AL, Velez E. 2014. Population structure and phylogeography reveal

pathways of colonization by a migratory marine reptile (*Chelonia mydas*) in the central and eastern Pacific. *Ecology and Evolution* 4(22): 4317-4331 DOI [10.1002/ece3.1269](https://doi.org/10.1002/ece3.1269).

Ekanayake EML, Kapurusinghe T, Saman MM, Rathnakumara DS, Samaraweera P, Rajakaruna RS. 2017. Genetic diversity of the Green Turtle (Testudines: Cheloniidae: *Chelonia mydas* (Linnaeus, 1758)) population nesting at Kosgoda Rookery, Sri Lanka. *Journal of Threatened Taxa* 9(6) DOI [10.11609/jott.2797.9.6.10261-10268](https://doi.org/10.11609/jott.2797.9.6.10261-10268).

Encalada SE, Lahanas PN, Bjorndal KA, Bolten AB, Miyamoto MM, Bowen BW. 1996.

Phylogeography and population structure of the Atlantic and Mediterranean green turtle *Chelonia mydas*: a mitochondrial DNA control region sequence assessment. *Molecular Ecology* 5(4):473-483 DOI [10.1111%2Fj.1365-294X.1996.tb00340.x](https://doi.org/10.1111%2Fj.1365-294X.1996.tb00340.x).

Frankham R. 2005. Genetics and extinction. *Biological Conservation* 126(2): 131-140 DOI [10.1016/j.biocon.2005.05.002](https://doi.org/10.1016/j.biocon.2005.05.002).

Formia A, Godley BJ, Dantaine JF, Bruford MW. 2006. Mitochondrial DNA diversity and phylogeography in African green turtles (*Chelonia mydas*). *Conservation Genetics* 7:353-369 DOI [10.1007/s10592-005-9047-z](https://doi.org/10.1007/s10592-005-9047-z).

Formia A, Broderick AC, Glen F, Godley BJ, Hays GC, Bruford MW. 2007. Genetic composition of the Ascension Island green turtle rookery based on mitochondrial DNA: implications for sampling and diversity. *Endangered Species Research* 3:145-158 DOI [10.3354/esr003145](https://doi.org/10.3354/esr003145).

Garzón-Ferreira J and Cano M. 1991. Tipos, Distribución, Extensión y Estado de Conservación de los Ecosistemas Marinos Costeros del Parque Nacional Natural Tayrona. Santa Marta: INVEMAR, 82.

Gaillard D, Yeh FC, Lin L, Chen HQ, Zhang T, Luo SJ, Shi HT. 2021. Lost at sea: determining geographic origins of illegally traded green sea turtles (*Chelonia mydas*) rescued on Hainan Island, China. *Wildlife Research* 48(1): 55-63 DOI 10.1071/WR19127.

Guo YS, Wang ZD, Liu CW. 2009. MtDNA Sequence Analysis of Green Sea Turtle (*Chelonia mydas*) of the South China Sea. *Journal of Guangdong Ocean University* 29(1): 6-9.

Gilpin ME, and Soulé ME. 1986. Conservation Biology: The Science of Scarcity and Diversity. *Ecological Restoration* 5(1): 23-24 DOI 10.3368/er.5.1.23.

Grant W, Bowen B. 1998. Shallow population histories in deep evolutionary lineages of marine fishes: Insights from sardines and anchovies and lessons for conservation. *Journal of Heredity* 89(5): 415-426 DOI 10.1093/jhered/89.5.415.

Hirth HF. 1997. Synopsis of the biological data on the green sea turtle, *Chelonia mydas* (Linnaeus 1758). *US Fish and Wildlife Service, Biol Rep* 97(1) <https://www.fao.org/3/c3466e/c3466e.pdf>.

Hamann M, Godfrey MH, Seminoff JA, Arthur K, Barata PCR, Bjoerndal KA, Bolten AB, Broderick AC, Campbell LM, Carreras C, Casale P, Chaloupka M, Chan SKF, Coyne MS, Crowder LB, Diez CE, Dutton PH, Epperly SP, FitzSimmons NN, Formia A, Girondot M, Hays GC, Cheng IS, Kaska Y, Lewison R, Mortimer JA, Nichols WJ, Reina RD, Shanker K, Spotila JR, Tomás J, Wallace BP, Work TM, Zbinden J, Godley BJ. 2010. Global research priorities for sea turtles: informing management and conservation in the 21st century. *Endanger Species Research* 11(3): 245-269 DOI 10.3354/esr00279.

Hughes AR, Inouye BD, Johnson MT, Underwood N, Vellend M. 2008. Ecological consequences of genetic diversity. *Ecology Letters* 11(6): 609-623

DOI 10.1111/j.1461-0248.2008.01179.x.

Hamabata T, Kamezaki N, Hikida T. 2014. Genetic structure of green turtle (*Chelonia mydas*)

peripheral populations nesting in the northwestern Pacific rookeries: evidence for northern refugia and postglacial colonization. *Marine Biology* 161(3): 495-507 DOI 10.1007/s00227-013-2352-z.

Jiang HS, Song XJ, Feng JN, Liu SY, Guan WP, Chen HF, Ouyang ZZ. 2000. Status of green sea turtles and their conservation in China. *Proceedings of the Fourth Asian Herpetological Conference* 1- 8.

Jia YY, Wang J, Balazs GH, Liu M. 2019. Nest productivity for green sea turtles (*Chelonia mydas*) at Qilianyu of Xuande Islands, South China Sea, P.R. China: Preliminary findings. *Chelonian Conservation and Biology* 18(1): 116-120 DOI 10.2744/CCB-1307.1.

Jian L, Guo R, Zheng XB, Shi HT, Wang JC. 2021. Trace elements in green turtle eggshells and coral sand sediments from the Xisha (Paracel) Islands, South China Sea. *Marine Pollution Bulletin* 164 DOI 10.1016/j.marpolbul.2021.112036.

Jensen MP, Bell I, Limpus CJ, Hamann M, Ambar S, Whap T, David C, FitzSimmons NN. 2016. Spatial and temporal genetic variation among size classes of green sea turtles (*Chelonia mydas*) provides information on oceanic dispersal and population dynamics. *Marine Ecology Progress Series* 543: 241-256 DOI 10.3354/meps11521.

Jensen MP, FitzSimmons NN, Bourjea J, Hamabata T , Reece J, Dutton PH. 2019. The evolutionary history and global phylogeography of the green sea turtle (*Chelonia mydas*). *Journal of Biogeography* 46(5): 860-870 DOI 10.1111/jbi.13483.

Lam T, Xu L , Takahashi S, Burgess EA. 2011. Market Forces: An Examination of Marine

Turtle Trade in China and Japan (TRAFFIC East Asia)

http://www.trafficchina.org/sites/default/files/traffic_species_reptiles_0.pdf.

Lacy RC. 1997. Importance of genetic variation to the viability of mammalian populations.

Journal of Mammalogy 78(2): 320-335 DOI 10.2307/1382885.

Lin L, Li SH, Chen M, Parham JF, Shi HT. 2021. Sea turtle demand in China threatens the

survival of wild populations. *iScience* 24(6): 102517 DOI 10.1016/j.isci.2021.102517.

Leroux RA, Dutton PH, Abreu-Grobois FA, Lagueux CJ, Campbell CL, Delcroix E,

Chevalier J, Horrocks Julia A, Hillis-Starr Z, Troëng S, Harrison E, Stapleton S. 2012.

Re-examination of population structure and phylogeography of hawksbill turtles in the wider

Caribbean using longer mtDNA sequences. *Journal of Heredity* 103(6): 806-820

DOI 10.1093/jhered/ess055.

Ng CKY, Dutton PH, Gu HX, Li TH, Ye MB, Xia ZR , Zhang FY, Duan JX, Hsu CK, Balazs

GH, Murphy MB. 2017. Regional conservation implications of green turtle (*Chelonia*

mydas) genetic stock composition in China. *Chelonian Conservation and Biology* 16(2):

139-150 DOI 10.2744/ccb-1253.1.

Mou JF, Tao CH, Ding XH, Wu FX, Miao X, Wang XY, Zhu Q. 2013. Investigations on the

distribution of sea turtle species in the coastal waters of China. *Journal of Applied*

Oceanography 32(2): 238-242, 294.

Marshall DR, Brown ADH. 1975. Optimum sampling strategies in genetic observation//Frankel

OH, Hawkes JG. Group genetic resources for today and tomorrow. *Cambridge: Cambridge*

university press 53-80.

Nishizawa H, Okuyama J, Kobayashi M, Abe O, Arai N. 2011. Population genetic structure and implications for natal philopatry of nesting green turtles *Chelonia mydas* in the Yaeyama Islands, Japan. *Endangered Species Research* 14: 141-148 DOI 10.3354/esr00355.

Naro-Maciel E, M Le, NN FitzSimmons, Amato G. 2008. Evolutionary relationships of marine turtles: a molecular phylogeny based on nuclear and mitochondrial genes. *Molecular Phylogenetics and Evolution* 49(2): 659-662 DOI 10.1016/j.ympev.2008.08.004.

Song JH, Lin BA, Jia YY, Dutton PH, Kang B, Balazs GH, Liu Min. 2022. New management unit for conservation of the Endangered green sea turtle (*Chelonia mydas*) at the Xisha (Paracel) Islands, South China Sea. *Endangered Species Research* 47: 145-154 DOI 10.3354/esr01172.

Stewart KR, Dutton PH. 2014. Breeding sex ratios in adult leatherback turtles (*Dermochelys coriacea*) may compensate for female-biased hatchling sex ratios. *PLoS One* 9(2): e88138 DOI 10.1371/journal.pone.0088138.

Stiles D. 2008. An Assessment of the Marine Turtle Products Trade in Vietnam. A TRAFFIC Southeast Asia Report. TRAFFIC, Southeast Asia, Petaling Jaya, Selangor, Malaysia.

Trifinopoulos J, Nguyen LT, von Haeseler A, Minh BQ. 2016. W-IQ-TREE: a fast online phylogenetic tool for maximum likelihood analysis. *Nucleic Acids Research* 44(W1): W232-W235 DOI 10.1093/nar/gkw256.

Vargas SM, Jensen MP, Ho SYW, Mobaraki A, Broderick D, Mortimer JA, Whiting SD, Miller J, Prince RIT, Bell IP, Hoenner X, Limpus CJ, Santos Fabrício R, FitzSimmons NN. 2016. Phylogeography, Genetic Diversity, and Management Units of Hawksbill Turtles in the Indo-Pacific. *The Journal of heredity* 107(3): 199-213 DOI 10.1093/jhered/esv091.

Wabnitz CCC, Lam VWY, Reygondeau G, Teh LCL, Al-Abdulrazzak D, Khalfallah M, Pauly D, Palomares MLD, Zeller D, Cheung WWL. 2018. Climate change impacts on marine biodiversity, fisheries and society in the Arabian Gulf. *PLoS One* 13 (5): e0194537 DOI 10.1371/journal.pone.0194537.

Wei WZ. 2016. The study on Conservation Biology of green sea turtle (*Chelonia mydas*) from South China Sea. East China Normal University.

Wang YM. 1993. The present situation and prospect of the investigation and conservation of sea turtle resources in the South China Sea. *Chinese Journal of Ecology* 12(6):60-61.

Wang J, Guo R, Yang YXY, Zhang Y, Jia YY, Liu M, Jin JM, Ji L, Zhang JQ. 2019. Threatened status and Recommendations on Ses Turtle Conservation in china. *Chinese Journal of Wildlife* 40(04):1070-1082.

Yang WJ. 2015. Analysis on genetic diversity of green sea turtle (*Chelonia mydas*) from South China Sea. Shandong University.

Yeh FC, Lin L, Zhang T, Green R, Martin F, Shi HT. 2021. Advancing sea turtle conservation in the South China Sea via U.S.-China diplomacy. *Environmental Progress & Sustainable Energy* 40(5) : e13643 DOI: 10.1002/ep.13643.

Zhang T, Lin L, Jian L, Wang Z, Kong L, Wang JC, Shi HT. 2020. Investigation of beach debris at nesting ground of green sea turtles (*Chelonia mydas*) at Qilianyu, Xisha (Paracel) Islands. *Journal of Ecology* 39(7): 2408-2415 DOI 10.13292/j.1000-4890.202007.035.

Zhang T, Lin L, Li DQ, Wu SN, Kong L, Wang JC, Shi HT. 2021a. The microplastic pollution in beaches that served as historical nesting grounds for green turtles on Hainan Island, China. *Marine Pollution Bulletin* 173(8):113069 DOI 10.1016/j.marpolbul.2021.113069.

Zhang T, Lin L, Gaillard D, Chen F, Chen H, Li M, Wu S, Wang Z, Shi HT. 2021b. Rapid authenticity testing of artificially bred green turtles (*Chelonia mydas*) using microsatellite and mitochondrial DNA markers. *PeerJ* 9:e12410 DOI [10.7717/peerj.12410](https://doi.org/10.7717/peerj.12410).

Zhang T, Lin L, Li DQ, Wang JC, Lin YT, Li R, Wu SN, Shi HT. 2022. Microplastic pollution at Qilianyu, the largest green turtle nesting grounds in the northern South China Sea. *PeerJ* 10:e13536 DOI [10.7717/peerj.13536](https://doi.org/10.7717/peerj.13536).